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Articles

Review of potentially harmful cyanobacteria

Revisión de cianobacterias potencialmente nocivas

Cesar Alejandro Zamora-Barrios¹, ORCID: <https://orcid.org/0000-0002-9536-440X>

Nandini Sarma², ORCID: <https://orcid.org/0000-0001-5614-6234>

S.S.S. Sarma³, ORCID: <https://orcid.org/0000-0003-2820-1579>

¹Aquatic Zoology Laboratory, Research and Postgraduate Division, Universidad Nacional Autónoma de México, Campus Iztacala, Los Reyes, Tlalnepantla, State of Mexico, Mexico, zamoracesaralejandro@hotmail.com

²Aquatic Zoology Laboratory, Research and Postgraduate Division, Universidad Nacional Autónoma de México, Campus Iztacala, Los Reyes, Tlalnepantla, State of Mexico, Mexico, nandini@unam.mx

³Aquatic Zoology Laboratory, Research and Postgraduate Division, Universidad Nacional Autónoma de México, Campus Iztacala, Los Reyes, Tlalnepantla, State of Mexico, Mexico, sarma@unam.mx

Corresponding author: Nandini Sarma, nandini@unam.mx



Abstract

A bibliographic review was developed by consulting various sources of information (articles, books, abstracts, etc.), obtained from the Web of Science database, Scopus, biological abstracts, etc. They describe the most relevant studies of the last three decades chronologically, from historical researches, as well as current topics under various sub-themes, critically analyzing nearly 200 articles with the aim to expose simply, but in a straightforward way, the general characteristics of cyanobacteria, the main conditions that favour the formation and persistence of blooms or "CyanoHABs"; negative implications on water resources due to cyanotoxin production, emphasizing the reference limits established for the hepatotoxin Microcystin-LR in water for human consumption, in recreational systems and food products; methodologies developed for monitoring toxic strains and a summary of the research published in Mexico on cyanobacteria and their toxins. Finally, some control procedures used in the remediation of systems with cyanobacterial blooms are discussed.

Keywords: Cyanobacteria, cyanohabs, cyanotoxins, microcystins.

Resumen

Se desarrolló una revisión bibliográfica a través de la consulta de diversas fuentes de información (artículos, libros, resúmenes, etc.) obtenidas de diversas bases de datos como *Web of Science*, *Scopus* y *Biological abstracts*, entre otras. Se describen de forma cronológica los estudios más relevantes de las últimas tres décadas, partiendo de investigaciones históricas, así como tópicos actuales bajo diversos subtemas; se



analizaron críticamente cerca de 200 de artículos con el objetivo de exponer de manera sencilla, pero explícita, las características generales de las cianobacterias; las principales condiciones que favorecen la formación y persistencia de los florecimientos o “CianoFANs”; las implicaciones negativas sobre los recursos hídricos debido a la producción de cianotoxinas, con énfasis en los límites de referencia establecidos para la hepatotoxina Microcistina-LR en agua de consumo humano, en sistemas de uso recreativo y en productos alimenticios; las metodologías desarrolladas para el monitoreo de cepas tóxicas, y un resumen de las investigaciones publicadas en México sobre cianobacterias y sus toxinas. Por último, se discuten algunos procedimientos de control usados en remediación de sistemas con proliferación de cianobacterias.

Palabras clave: cianobacteria, cianofan, cianotoxinas, microcistinas.

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Cyanobacteria: general features

Cyanobacteria or cyanoprokaryote are the first oxygenic photoautotrophs (Komárek, 2006). According to the classification based on phylogenetic



relationships, they belong to the eubacterial domain (Maddison, Schulz, & Maddison, 2007). Their age is estimated at more than 3.5 billion years through fossil records (microbialites: organosedimentary deposits formed by accretion, precipitation, and union of mineral materials) and geological data obtained from carbon isotopic samples consistent with CO₂ fixation through the Rubisco enzyme (Castle & Rodgers, 2009; Schopf, 2012). They are responsible for altering the primitive biosphere and transforming it into an oxidizing one during the Precambrian (Schirrmeister, de Vos, Antonelli, & Bagheri, 2013; Pla-García & Menor-Salván, 2017; Moss, 2018). Cyanobacteria have colonized most ecological niches due to their long evolutionary history. They are found from the poles to the tropics in aquatic and terrestrial systems, tolerating inhospitable environmental circumstances of low and high alkalinity and predominating in geothermal microbial communities (Steinberg, Schäfer, & Beisker, 1998; Miller & Castenholz, 2000). However, cyanobacterial taxa require different ecological strategies to grow in each specific habitat. Most species have ecological niches restricted by their morphological and physiological traits (Komárek, 1995; Mateo, Leganés, Perona, Loza, & Fernández-Piñas, 2015).

Among the prokaryotes and from the study of morphological characters (traditional taxonomy), cyanobacteria comprise about 24 % of them, with about 3 000 species arranged in 150 genera (Guiry, 2012; Nabout, Da-Silva-Rocha, Carneiro, & Sant'Anna, 2013). Based on a polyphasic approach that integrates phenotypic, ecophysiological and phylogenetic information, reveals that cyanobacteria are a widely diversified group. To understand their evolutionary relationships and generate a classification that does not underestimate or overestimate the

size of the populations, they have been divided into eight orders: Gloeobacterales, Synechococcales, Spirulinales, Chroococcales, Pleurocapsales, Oscillatoriales, Chroococcidiopsidales, Nostocales (Komárek, 2010; Komárek, 2014). However, using mathematical logistic models based on the species recognized so far, it has been estimated that this group must contain about 3 580 species which have not yet been described (Nabout *et al.*, 2013).

Cyanobacteria have a wide range of morphological variations, unicellular organisms from 2 to 20 μm and multicellular filaments or colonies exceeding 500 μm (Whitton, 2002; Codd, Lindsay, Young, Morrison, & Metcalf, 2005b). They reproduce asexually by binary or multiple fission; however, some species can produce endospores and exospores (exocytes, baeocytes and hormocytes). Another reproductive strategy is the fragmentation of colonies or filamentous trichomes in the form of hormogonia (short trichome fragments separated by necridial zones formed by programmed cell death) (Whitton & Potts, 2012). Cell division occurs in one, two, or three perpendicular or irregular planes, resulting in a variety of three-dimensional configurations (Mazouni, Domain, Cassier-Chauvat, & Chauvat, 2004; Flores & Herrero, 2014). Furthermore, some filamentous cyanobacteria have different types of branching, which gives them a distinctive morphology used for taxonomic classification (Cirés, 2012). Unlike photosynthetic eukaryotes, they do not have a defined nucleus and lack plastids and organelles. Likewise, their cell wall is Gram-negative, free of teichoic acids, formed by three main layers composed of peptidoglycan; although some genera present a fourth protein layer, called the superficial layer or S-layer (Rippka,

Deruelles, Waterbury, Herdman, & Stanier, 1979; Hoiczky & Hansel, 2000).

The photosynthetic pigment composition of cyanobacteria is variable, as they synthesize a wide range of accessory pigments, such as phycocyanin (maximum absorbance = 620 nm), phycoerythrin (maximum absorbance = 565 nm), allophycocyanin (maximum absorbance = 650 nm), and various carotenoids, resulting in a wide range of colorations, including red, brown, and pink; However, one of the transcendental features is their distinctive blue-green (cyan) coloration, which is due to the ubiquitous presence of the pigment chlorophyll-a hidden by phycobiliproteins (Huisman *et al.*, 2018).

Cyanobacteria play an essential role in energy flow, transforming inorganic matter into organic matter in biomass. At the same time, they play an important role in both nitrogen and carbon biogeochemical cycles (Bellinger & Sigee, 2015). The growth of primary producers depends on a quota of essential inorganic nutrients; these nutritional requirements have a stoichiometric relation known as the Redfield ratio, an average proportion of the atomic composition of phytoplankton biomass, with a relatively constant ratio of 106 carbon to 16 nitrogen and 1 phosphorus atoms (Redfield, 1958; Ptacnik, Andersen, & Tamminen, 2010). Nitrogen and phosphorus are vital nutrients for primary producers, structuring the synthesis of cell membranes, DNA (deoxyribonucleic acid), amino acids, proteins, and as part of cell metabolism (Reynolds, 2006).

Cyanobacteria actively incorporate a wide range of nitrogen molecules such as ammonium (NH_4^+), nitrate (NO_3^-), nitrite (NO_2^-), urea (CH_4NO_2), etc. Depending on the nitrogen source, assimilation requires

additional enzymatic reductions (nitrate reductase and nitrite reductase). Ammonium is the least expensive nitrogen source; for glutamate synthesis, one molecule of NADPH (Nicotinamide-Adenine-Dinucleotide-Phosphate) and ATP (Adenosine triphosphate) is required (Flores & Herrero, 2005). The main phosphorus supply for cyanobacteria is orthophosphates (H_3PO_4), but they also use other organic forms (Whitton & Potts, 2012). Low N:P (nitrogen: phosphorus) ratios have been considered as one of the factors influencing the success of cyanobacteria, and conversely, a high ratio (50:1) favors their dominance in epicontinental systems (Hyenstrand, Blomqvist, & Pettersson, 1998; Bulgakov & Levich, 1999).

CyanoHABs: Factors and traits that drive your success

Cyanobacteria significantly accelerate the average population biomass in a relatively short time, generating massive growth phenomena ($> 10^4$ cels m l^{-1}), known as harmful algal blooms "CyanoHABs" (Carmichael, 2001; Codd, Morrison, & Metcalf, 2005a; Ger, Hansson, & Lürding, 2014). A CyanoHAB is distinguished by the dominance of a few species, which account for more than 80 % of the total biomass among primary

producers (Mihaljević & Stević, 2011; Humbert & Fastner, 2017). CyanoHAB has a severe impact on the water quality of aquatic systems caused by the organic matter accumulation that decreases the vertical transparency, causing alterations in the submerged aquatic vegetation.

Despite not being toxic to humans, cyanobacteria produce volatile organoleptic compounds such as geosmin and 2-methylisoborneol, which interfere with the recreational function of aquatic systems or their use as drinking water reservoirs by altering public perceptions of hygiene and public health and causing significant economic losses in both sectors (Liato & Aider, 2017; Churro, Semedo-Aguiar, Silva, Pereira-Leal, & Leite, 2020).

Furthermore, due to photosynthesis and cellular respiration, dense blooms cause dramatic alterations in dissolved oxygen concentrations (Vermaas, 2001; Moss *et al.*, 2011). They coexist in microenvironments with heterotrophic bacteria immersed in the phycosphere (the region surrounding a cyanobacterial cell enriched with exuded organic molecules) (Steiner *et al.*, 2017), which during the senescence of a bloom promote hypoxia and anoxia periods due to degradation process, causing the death of benthic, nektonic and planktonic organisms, (fish, birds, zooplankton, phytoplankton, etc.), decreasing the biodiversity of aquatic ecosystems (Oberholster, Botha, & Cloete, 2006; Paerl & Otten, 2013).

Eutrophication increased atmospheric carbon dioxide (CO₂), and global climate change are just a few of the many environmental factors contributing to blooms frequency and intensity (Huisman *et al.*, 2018). CyanoHABs have been more intense in the last 50 years as the rate of anthropogenic eutrophication has increased due to industrialization and

urbanization processes (Paerl & Paul, 2012; Deshpande, Tremblay, Pienitz, & Vincent, 2014; Paerl, 2017).

Global warming stimulates the proliferation of cyanobacteria directly or indirectly, wreaking havoc on the functioning of aquatic systems and ecological structure (Visser, Ibelings, Bormans, & Huisman, 2016a; Huisman *et al.*, 2018).

Directly, global warming favors increased photosynthetic capacity and accelerated metabolic rate, presenting its maximum population growth rates above 25 °C Lüring, Eshetu, Faassen, Kosten, & Huszar, 2013; Paerl, 2014; Savadova *et al.*, 2018). Thermal increase stimulates the reintegration of cells or colonies of meroplanktonic genera (Cirés, Wörmer, Agha, & Quesada, 2013), incite early akinete inoculation, and increase of the germination percentage (Carey, Ibelings, Hoffmann, Hamilton, & Brookes, 2012; Cirés *et al.*, 2013; Silveira & Odebrecht, 2019). On the other hand, global warming promotes biogeographic range expansion, establishment, and dispersal of species with invasive potential such as *Raphidiopsis raciborskii* (Nostocales), *Planktothrix rubescens* (Oscillatoriales), and *Synechococcus capitatus* (Chroococcales) (Sinha *et al.*, 2012; Kokociński, Akçalan, Salmaso, Stoyneva-Gärtner, & Sukenik, 2017). Indirect effects of temperature involved higher stratification stability on the water column with vertical mixing reduction because of changes in water density (Joehnk *et al.*, 2008; Paerl *et al.*, 2020) and viscosity decrease, causing loss due to sedimentation of those primary producers without buoyancy regulating structures (Joehnk *et al.*, 2008; O'Neil, Davis, Burford, & Gobler, 2012).

In recent decades atmospheric CO₂ emissions have increased at a rate of 3 % per year, derived from the combustion of fossil fuels (O'Neil *et al.*, 2012). Due to their physiological flexibility, cyanobacteria have inorganic carbon concentration systems, which give them efficient CO₂ and HCO₃⁻ consumption (Huisman *et al.*, 2018). Nevertheless, since high photosynthetic rates generate higher demand for CO₂, cyanobacteria with gas vesicles have a competitive advantage because they directly intercept this ion from the atmosphere (Paerl & Huisman, 2009). An increase in CO₂ concentrations in the atmosphere probably favors CyanoHABs in eutrophic waters, especially those with the possibility of regulating their vertical buoyancy along the water column and combining carbon concentration systems (Ji, Verspagen, Stomp, & Huisman, 2017).

Hydrological alterations are another indicator of climate change. Prolonged and heavy precipitation enhance runoff and the discharge of nutrient-rich groundwater into drainage basins, which occurs naturally because of erosion. This inflow of large volumes of water during storms favors sediment removal, subsequent release of nutrients into the water column, and reintegration of cyanobacterial cells in dormancy (Cirés *et al.*, 2013; Havens, East, & Beaver, 2016). However, if storms are followed by periods of drought and long water residence stages, populations of slow growing and persistent primary producers, such as cyanobacteria, are favored (Moss *et al.*, 2011; Reichwaldt & Ghadouani, 2012; Chapra *et al.*, 2017).

Factors intrinsic to cyanobacterial cells also contribute to their ecological success. Cyanobacteria tend to change their unicellular to multicellular morphology into large colonies or filaments, avoiding this

way their ingestion by size or exclusion by interference from feeding appendages of filter-feeding herbivores (Porter, 1973; DeMott, Gulati, & Van Donk, 2001). Several species are embedded into a colloidal element composed of mucilage which is an extracellular polymeric substance which acts a protective boundary between cells and the surrounding environment and plays a key role in intraspecific and interspecific interactions (Kehr & Dittmann, 2015).

Under phosphorus limiting conditions, cyanobacteria form polyphosphate granules, used as a reserve, which is enough to perform several cell divisions (Whitton & Potts, 2012). Also, cyanobacteria can transform atmospheric N₂ by different enzymatic processes (diazotrophy) during nitrogen limiting conditions. a procedure that requires a significant amount of energy (10 % of total cellular proteins and 16 ATP molecules) (Berman-Frank, Quigg, Finkel, Irwin, & Haramaty, 2007). Diverse filamentous cyanobacteria (mainly of the Nostocales genus) carry out fixation in morphologically modified microaerobic cells, known as heterocytes or diazocytes, where the enzyme nitrogenase is expressed (Bergman, Sandh, Lin, Larsson, & Carpenter, 2013).

Planktonic species can benefit when exploiting water systems with dense thermal stratification, regulating buoyancy through the metalimnion, or if the systems are unstable, they occupy the surface regions of the epilimnion, which are zones with better illumination and higher nutrient concentration. This is due to the production of intracellular gas vesicles, aerotopes, which are arranged in structures that are distributed parallel to one another (Walsby, Hayes, Boje, & Stal, 1997; Mur, Skulberg, & Utkilen, 1999; Kobos *et al.*, 2013).

Mechanisms behind buoyancy involve modulation of genetic expression of vesicles, their collapse by turgor pressure, and the action of ballast, such as intracellular carbohydrates the accumulation of intracellular carbohydrates produced during the respiration (Walsby, 1994; Wallace, Bailey, & Hamilton, 2000; Wörmer, Cirés, & Quesada, 2011). Larger colonies have higher ascent rates because the buoyancy rate is related to its diameter squared (Rabouille, Salençon, & Thébault, 2005; Zhao *et al.*, 2016). Field studies demonstrate that colonies larger than $> 300 \mu\text{m}$ resist turbulent mixing, whereas colonies smaller than $< 36 \mu\text{m}$ disperse uniformly and are susceptible to sedimentation (Chien, Wu, Chen, & Chou, 2013; Zhao *et al.*, 2016).

Bioactive metabolites: Cyanotoxins

Cyanobacteria can produce bioactive metabolites with varied chemical structures; more than 2 000 metabolic substances have been characterized so far (Carmichael *et al.*, 2001; Jones *et al.*, 2021). Due to their high complexity, these substances draw the attention of researchers for use in the field of biotechnology and are commercially exploited in the food, cosmetics, agrochemical, and pharmaceutical industries (Liu, Pohnert, & Wei, 2016).



Pharmaceutical research explores their use to combat against cancer, cardiac disorders, autoimmune and infectious diseases (Swain, Paidesetty, & Padhy, 2017). Compounds with anticancer activity derived from cyanobacteria include alkaloids, polyketides, terpenes, peptides, nucleosides, and carbohydrates. These include cytotoxic compounds targeting non-specific macromolecules expressed by cancer cells or metabolites that alter oncogenic signal transduction pathways (Nobili *et al.*, 2009; Mondal *et al.*, 2020).

Among the bioactive secondary metabolites, the so-called cyanotoxins stand out mainly because of their adverse effect on mammals, birds, fish, zooplankton, protozoa, bacteria, and their harmful impact on human health (Yadav, Sinha, Tyagi, & Kumar, 2011).

These substances are classified chemically as cyclic peptides (heptapeptides, such as Microcystins, and pentapeptides, such as Nodularin), alkaloids (anatoxins, saxitoxins, and cylindrospermopsin), and lipopolysaccharides (Whitton & Potts, 2012). They can also be categorized as cytotoxins or biotoxins if they affect an organ or tissue (hepatotoxins, neurotoxins, and irritant toxins) (Vasconcelos, 2001; Whitton & Potts 2012; Ibrahim, Khairy, & Ibrahim, 2012; Paerl, Otten, & Joyner, 2016; Huisman *et al.*, 2018) (Table 1).

Table 1. Scheme of cyanotoxins, characteristics and examples of the main producing genera (Adapted from Huisman *et al.*, 2018).

Cyanotoxin	Chemical Structure	Classification and Mechanisms of Action	Main Producing Genera
Microcystins	Cyclic heptapeptide	Hepatotoxins: Inhibition of eukaryotic protein phosphatases (PP1 and PP2A). Hepatic and renal damage, gastroenteritis, tumor promoter.	<i>Microcystis</i> <i>Dolichospermum</i> <i>Planktothrix</i> <i>Nostoc</i> <i>Anabaenopsis</i> <i>Leptolyngbya</i> <i>Phormidium</i> <i>Synechococcus</i>
Nodularins	Cyclic pentapeptide	Hepatotoxins: Inhibition of eukaryotic protein phosphatases. Promoters of cancerous tumors, liver damage.	<i>Nodularia</i>
Anatoxin-a	Bicyclic alkaloid	Neurotoxin: Agonist of nicotinic acetylcholine receptors at neuromuscular junctions. Blocker of postsynaptic neuromuscular depolarization. Loss of coordination, muscle spasms and respiratory failure.	<i>Anabaena</i> <i>Aphanizomenon</i> <i>Cuspidothrix</i> <i>Dolichospermum</i> <i>Planktothrix</i> <i>Oscillatoria</i> <i>Phormidium</i>
Anatoxin-a(s)	Alkaloid	Neurotoxin: Acetylcholinesterase inhibitor. Salivation, muscle spasms and respiratory failure.	<i>Dolichospermum</i>
Saxitoxins	Alkaloid	Neurotoxins: Blocks sodium ion channels of eukaryotic neurons. Promotes numbness, paralysis, and respiratory failure.	<i>Aphanizomenon</i> <i>Cuspidothrix</i> <i>Cylindrospermopsis</i> <i>Dolichospermum</i> <i>Lyngbya</i> <i>Planktothrix</i>

Cyanotoxin	Chemical Structure	Classification and Mechanisms of Action	Main Producing Genera
Cylindrospermopsin	Tricyclic guanidine alkaloid	Multiple toxicity. Neurotoxic, genotoxic and protein synthesis inhibitor potential.	<i>Cylindrospermopsis</i> <i>Umezakia</i> <i>Anabaena</i> <i>Oscillatoria</i> <i>Raphidiopsis</i>
BMAA	Beta-methylamino-L- alanine (non-encoded amino acids)	Neurotoxin: Excessive stimulation of glutamate receptors in neurons and erroneous insertion into proteins, changing their configuration. Linked to neurodegenerative diseases.	<i>Microcystis</i> <i>Nostoc</i> Possibly widespread among more cyanobacteria, but still no consensus.
LPS	Lipopolysaccharides	Dermatotoxins: Inflammation and promoters of cytokine production. Skin irritation, fever and gastrointestinal distress.	All cyanobacteria.

Over 50 % of the CyanoHAB-forming genera can produce one or more cyanotoxins (Codd *et al.*, 2005a; Codd *et al.*, 2005b). Cyanotoxins are endotoxins, and their release into the environment is regulated by cell membrane lysis, especially during CyanoHAB senescence (Chorus & Bartram, 1999). However, some physical or chemical agents promote its disruption and cyanotoxin release (Hoeger, Dietrich, & Hitzfeld, 2002; Dai *et al.*, 2016). During the exponential growth phase, 10-20 % of the concentration is released to the surrounding water (Watanabe, Tsuji, Watanabe, Harada, & Suzuki, 1992; Negri, Jones, Blackburn, Oshima, & Onodera, 1997; Rapala, Sivonen, Lyra, & Niemelä, 1997). In some cases, dissolved cyanotoxin ratios are higher than intracellular concentrations (Bumke-Vogt, Mailahn, & Chorus, 1999; Kinnear, Duivenvoorden, &

Fabbro, 2007). Once in the environment, they have variable persistence (e.g., the half-life of microcystins is up to 10 weeks at 40 °C and pH 10) because they can be biodegraded by bacteria, which can hydrolyze the molecule and use it as a carbon or nitrogen source (strains of *Sphingomonas*, *Sphingosinicella*, *Novosphingobium*, *Sphingopyxis*, and *Stenotrophomonas*), physicochemically decomposed (e.g., by photolysis) or absorbed by sedimentary particles (Edwards & Lawton, 2009; Harada & Tsuji, 1998; Brinkman & Bourne, 2013; Schmidt, Wilhelm, & Boyer, 2014).

Microcystins

Microcystins (MCs) are the most abundant cyanotoxins in both freshwater and brackish habitats around the world (Chorus & Bartram, 1999), with approximately 280 known congeners, which are distinguished by amino acids linked to the X and Z positions of the heptapeptide structure (seven amino acids in a ring formation with a single β -amino acid side chain (ADDA group) (Chorus & Bartram, 1999; Bouaïcha *et al.*, 2019). Environmental factors such as competition, predation, temperature, light intensity, and nutrition concentration influence the production of MCs (Merel *et al.*, 2013).

MCs are not synthesized by the ribosomal pathway; their gene expression depends on genomic locus coding *mcy* (*Microcystis* and *Anabaena*) arranged in two divergent transcription operons. Multifunctional enzyme complexes containing nonribosomal peptide synthetase (*NRPS*) and polyketide synthase (*PKS*) domains (Tillett *et al.*, 2000).

The *Microcystis* genus mostly produces this toxin (e.g., *Microcystis aeruginosa*, *Microcystis novacekii*, *Microcystis panniformis*, *Microcystis protocystis*). Nevertheless, because of horizontal gene transfer, about 23 planktonic and benthic taxa, including *Anabaena*, *Phormidium*, *Planktothrix*, *Nostoc*, *Limnothrix*, *Anabaenopsis*, *Aphanocapsa*, *Aphanizomenon*, *Cylindrospermopsis*, *Fischerella*, *Hapalosiphon*, *Lyngbya*, *Oscillatoria*, *Rivularia*, *Synechocystis* and *Synechococcus*, can synthesize (Rantala *et al.*, 2004; Rastogi, Madamwar, & Incharoensakdi, 2015; Catherine, Bernard, Spoof, & Bruno, 2017).

Exposure to MCs represents a high risk to humans. Once ingested, they cannot be hydrolyzed by peptidase enzymes in the stomach, and due to their high molecular weight (from 900 to 1 100 Da), they do not passively diffuse across cell membranes; therefore, they are actively transported to hepatocytes by organic anion transporter proteins of the biliary system (OATP1B1 and OATP1B3) (Popovic, Zaja, & Smital, 2010; Fontanillo & Köhn, 2018). MCs have an inhibitory effect on serine and threonine enzymes responsible for catalyzing phosphorylation. The toxicity is caused by the amino acid Adda from microcystin binding to the catalytic site of protein phosphatases (PP1, PP2A, PP2B, PP4, PP5, and PP6), which inhibits substrate dephosphorylation (Mackintosh, Beattie,

Klumpp, Cohen, & Codd, 1990; Swingle, Ni, & Honkanen, 2007), promoting cell necrosis or apoptosis, damage to DNA repair systems, gene expression and oxidative stress (Campos & Vasconcelos, 2010; Zanchett & Oliveira-Filho, 2013). Gastroenteritis, irritations, liver disorders, including necrosis and cancer, are the most common symptoms of MC poisoning, which can lead to death (Fontanillo & Köhn, 2018).

Guiding levels and guidelines

Due to its high degree of toxicity and frequency around the world (McLellan & Manderville, 2017), different environmental authorities, including the World Health Organization (WHO), established a reference value of MCs of $1 \mu\text{g l}^{-1}$ for drinking water sources of the LR variant (Chorus & Bartram, 1999). The same level was set in Mexico under the draft standard PROY-NOM-127-SSA1-2017. Furthermore, to reduce the health risks caused by toxin-producing cyanobacteria, a guide of reference values for recreational waters has been established, based on cell density and microcystin concentration. Recommended concentrations representing low risk are $4\text{--}10 \mu\text{g l}^{-1}$ or $< 20\,000$ cyanobacterial cells ml^{-1} , $< 20 \mu\text{g l}^{-1}$, while a concentration of $20 \mu\text{g l}^{-1}$ or between $20\,000$ and $100\,000$ cyanobacterial cells ml^{-1} is considered moderate risk, and a concentration above $20 \mu\text{g l}^{-1}$ or $> 100\,000$ cyanobacterial cells ml^{-1}

carries a high risk of adverse health effects (Table 2) (Chorus & Bartram, 1999; Churro, Dias, & Valério, 2012; Paerl & Otten, 2013).

Table 2. Guidelines for safe practice in recreational water management (modified from WHO, 2003; Chorus & Bartram, 1999; Churro *et al.*, 2012).

Guideline or risk level	Orientation status	Human health hazards	Recommended actions
Low	<20 000 cyanobacterial cells per ml ⁻¹ < 10 µg l ⁻¹ chlorophyll-a with cyanobacterial dominance <2.5 mm ³ l ⁻¹ cyanobacterial biomass 4-10 µg l ⁻¹ of microcystins-LR	Low probability of adverse health effects	Continue monitoring
Moderate	20 000-100 000 cyanobacterial cells per ml ⁻¹ 10-50 µg l ⁻¹ chlorophyll-a with cyanobacterial dominance 2.5-12.5 mm ³ l ⁻¹ cyanobacterial biomass 10-20 µg l ⁻¹ of microcystin-LR	Short-term adverse health effects, such as skin irritations and gastrointestinal diseases	Add signs to indicate MODERATE warning level, indicating increased health risk when swimming or performing activities involving contact with water
High	>100 000 cyanobacterial cells per ml ⁻¹ >50 µg l ⁻¹ chlorophyll-a with cyanobacterial dominance >12.5 mm ³ l ⁻¹ cyanobacterial biomass >20 µg l ⁻¹ of microcystin-LR	Short-term adverse effects on human health, such as skin irritations or gastrointestinal illness following accidental contact or ingestion Severe acute poisoning is possible in the worst cases of ingestion	Immediate action to prevent contact with the bloom Add signs to indicate HIGH alert level, warning of danger when swimming or other activities involving contact with water Inform relevant authorities. Follow-up public health investigation

Most reports of human poisoning by MCs assess direct exposure through drinking water supply (Liu *et al.*, 2011; Tian *et al.*, 2013); clinical failures in patients undergoing renal hemodialysis with diluted cyanotoxins (Jochimsen *et al.*, 1998; Azevedo *et al.*, 2002); cyanobacterial cell contact and ingestion during recreational activities (Pilotto *et al.*, 1997; Stewart, Schluter, & Shaw, 2006), and consumption of oral dietary supplements from cultures contaminated with toxin-producing strains of cyanobacteria (Saker, Welker, & Vasconcelos, 2007; Costa *et al.*, 2018). However, vectors accumulating these secondary metabolites provide indirect routes of cyanotoxin uptake.

The main routes through ingestion of contaminated products include consumption of agricultural products irrigated with treated water containing cyanotoxin (Drobac *et al.*, 2013) and eating seafood from sites where toxin-producing cyanobacteria are prevalent (Ibelings & Chorus, 2007; Ferrão-Filho & Kozłowsky-Suzuki, 2011), which represents a recurrent problem in aquaculture ponds worldwide (Mohamed, Carmichael, & Hussein, 2003; Drobac *et al.*, 2016). Fish ingest and accumulate cyanotoxins directly through water intake or consumption cyanoprokaryote cells as part of their normal diet (omnivores and phytoplanktivores) (Ibelings *et al.*, 2005). The indirect route develops when epithelial cells are exposed with the diluted fraction (skin and gills) or by accumulation and transfer through foodweb (Zamora-Barrios, Nandini, & Sarma, 2019).

Under laboratory conditions, MCs are molecules resistant to denaturation by boiling and preserving their stability at temperatures above 300 °C (Wannemacher, 1989). Cooking food contaminated with

MCs does not favor their degradation; on the contrary, it increases their bioavailability by breaking the covalent bonds with the catalytic subunits of phosphatases, increasing the risk during the eating of broths or seafood soups (Morais, Augusto, Carvalho, Vale, & Vasconcelos, 2008; Zhang, Xie, & Chen, 2010). Additionally, to toxic components, fish fed with cyanobacteria present a deficit of protein and lipid nutrients as well as a bad taste due to the accumulation of organoleptic compounds (Liang, Zhou, Zhang, Qiao, & Zhang, 2015).

WHO has also established a provisional guideline level for the tolerable daily intake of contaminated foods with microcystin, which was set at $0.04 \mu\text{g}^{-1} \text{kg}^{-1} \text{d}^{-1}$, and is derived from the no-observed adverse effect level (NOAEL) of $40 \mu\text{g kg}^{-1}$ of microcystin-LR, based on a histopathological study in mouse, and supported by the lowest observed adverse effect level (LOAEL) obtained from a chronic study of the effect of crude extracts of *Microcystis* in the drinking water of pigs; however, due to interspecies variability an uncertainty value of 1 500 is applied (Fawell, James, & James, 1994; Chorus & Bartram, 1999).

Detection, identification, and evaluation methods

Due to the worldwide interest in solving public health problems and early detection of water safety difficulties, several methodologies have been

developed to determine, identify, and evaluate cyanotoxin concentrations, as well as to know the risk potential of cyanobacteria (Codd *et al.*, 2001).

High performance liquid chromatography (HPLC) is the most extensively used chemical method along with UV-Visible, PDA (photodiode array detection) and FLD (fluorescence detection) detectors. These methods are based on the separation of each component of a mixture, according to different types of chemical interactions, such as polarity (extraction) and unequivocally identifying the analyte (Berry, 2013). Enzymatic approaches such as inhibiting serine/threonine phosphatases (microcystins and nodularin), cholinesterases (anatoxins), or binding to specific receptors were also developed (Vogiaz *et al.*, 2019).

The immunological ELISA test is one of the most sensitive and is widely used to identify and quantify cyanotoxins in raw or drinking water, as well as in complex matrices including animal or plant tissues, sediments, urine, and human plasma (Foss & Aubel 2013; Moreira, Ramos, Azevedo, & Vasconcelos, 2014). Monoclonal kits are based on competition between the antigen diluted in the sample (specific cyanotoxin) and the peroxidase enzyme labeled (H_2O_2 ; conjugated enzyme) for the limited binding sites of a specific antibody (Ueno *et al.*, 1996; Adamovský *et al.*, 2007). The USEPA (United States Environmental Protection Agency) (USEPA, 2015) recommends the ELISA immunoassay kit as the preferred analytical screen for quantification of these secondary metabolites in aquatic systems because of the rapidity with which results can be acquired (2h).

On the other hand, bioassays provide information on effects on biological entities such as populations, organisms, cell cultures, purified molecules, specific genetic effects, and physiological conditions (depending on the degree of sensitivity to be assessed) (Meriluoto, Metcalf, & Codd, 2017). Thus, studies have been conducted on crustaceans, protozoa, insects, rotifers, cnidaria, nematodes, oligochaetes, plants, etc. (Maršálek & Bláha, 2004).

Cyanobacteria toxicity has been evaluated using monoclonal strain cultures (Hughes, Gorham, & Zehnder, 1958; Watanabe & Oishi, 1985; Zamora-Barrios, Nandini, & Sarma, 2015); purified cyanotoxins (DeMott, Zhang, & Carmichael, 1991; Ghadouani, Pinel-Alloul, Plath, Codd, & Lampert, 2004; Huang, Xi, Xu, & Wen, 2012), and crude extracts of cyanobacterial consortia (Zamora-Barrios, Nandini, & Sarma, 2017; Pawlik-Skowrońska, Toporowska, & Mazur-Marzec, 2019; Janssen, 2019); it is suitable to evaluate their toxicological effect through bioassays, when there are consortia of strains producing several biotoxins or if blooms are dominated by poorly studied species that might contain unknown metabolites with variable toxicity (Bláha *et al.*, 2017).

Molecular detection techniques are based on amplifying specific regions, using primers of genomic DNA and ribosomal RNA by polymerase chain reaction (PCR), which allows the comparison of copies of nucleotide sequences of coding genetic regions. One universally used marker for taxonomic identification in cyanobacteria is the small-subunit ribosomal called 16S rRNA (Nübel, Garcia-Pichel, & Muyzer, 1997; Albrecht, Pröschold, & Schumann, 2017). Nevertheless, due to the limited variability between genera, amplification of genes encoding enzymes

involved in metabolic activities or intergenic space usage, such as the phycocyanin operon (*PC-IGS*) or dinitrogenase reductase (*nifH* and *nifD*), is occasionally necessary (Teneva, Dzhambazov, Mladenov, & Schirmer, 2005; Hartmann & Barnum, 2010). Toxin-producing strains are detected by amplification of genes involved in their biosynthesis (e.g., the *mcy* group for microcystin and *cyr* for cylindrospermopsin). This absence provides a qualitative tool to discriminate potentially harmful strains from non-toxic strains ((Baker *et al.*, 2013; Salmaso *et al.*, 2016). Moreover, fluorescent markers (qPCR) have been developed to provide a quantitative real-time estimation of the number of cells containing encoding genes or transcript levels implicated in cyanotoxin biosynthesis (Humbert, 2017; Kurmayer, Sivonen, Wilmotte, & Salmaso, 2017).

Main evaluations in Mexico

CyanoHABs have been recorded in Mexico for more than three decades, and it is recognized that genera with toxicological potential, such as *Microcystis*, *Planktothrix*, and *Anabaenopsis*, have dominated the biomass of primary producers (Ortega, 1984; Alcocer, Kato, Robles, & Vilaclara, 1988); however, the first reports of cyanotoxin detection in Mexican water bodies were published 20 years later (Berry & Lind, 2010a; Vasconcelos *et al.*, 2010). Cyanotoxin presence has focused on detecting microcystins, with concentrations ranging from 0.03 to more than 70 g l⁻¹. The



occurrence of cylindrospermopsin and saxitoxins accumulated in the tissues and organs of fish and the snail *Pomacea patula catemacensis* has also been confirmed, all of which are commercially important in Lake Catemaco, Veracruz (Berry & Lind, 2010b; Berry, Jaja-Chimedza, Dávalos-Lind, & Lind, 2012). A recent study (Zamora-Barrios *et al.*, 2019) shows that fish ingested whole as *Chirostoma* sp. represent a higher health risk than tilapia, where only the muscles are consumed. In the latter, organs such as the liver and digestive tract, where most cyanotoxins bioaccumulate, are discarded. Each of the epicontinental systems evaluated provides ecosystem services to the surrounding human populations.

Most evaluations in Mexican laboratories focus on experiments with cyanobacterial strains from natural systems to assess their potential use in biomanipulation, considering the ability of the organisms to consume the cells (Nandini, Sarma, & Ramírez-García, 2000; Fernández, Nandini, Sarma, & Castellanos-Páez, 2014; Figueroa-Sánchez, Nandini, Castellanos-Páez, & Sarma, 2019). The synergistic impact of environmental stressors (predation and competition) experienced by filter-feeding species commonly found in water bodies with the presence of CyanoHABs; in addition to the evaluation of environmental factors (herbivory, nutrient concentration, temperature, and light intensity) that trigger gene expression and microcystin synthesis (Pineda-Mendoza, Zúñiga, & Martínez-Jerónimo, 2014; Pérez-Morales, Sarma, & Nandini, 2015; Pineda-Mendoza, Zúñiga, & Martínez-Jerónimo, 2016), also the antibacterial effect of bioactive compounds isolated from cyanobacteria (Gutiérrez, Flores, Solís, & Jimenez, 2008), the deleterious effects of exudates, crude extracts of monoclonal strains, and mixtures of

cyanobacteria from natural systems on zooplankton organisms (Arzate-Cárdenas, Olvera-Ramírez, & Martínez-Jeronimo, 2010; Olvera-Ramírez, Centeno-Ramos, & Martínez-Jerónimo, 2010; Pineda-Mendoza, Olvera-Ramírez, & Martínez-Jerónimo, 2012; Zamora-Barrios *et al.*, 2015; Zamora-Barrios *et al.*, 2017; Nandini, Sánchez-Zamora, & Sarma, 2019; Nandini, Zamora-Barrios, & Sarma, 2020) or by mixed diets, composed of different proportions of *Microcystis aeruginosa* and green algae (Alva-Martínez, Sarma, & Nandini, 2007a; Alva-Martínez, Sarma, & Nandini, 2007b; Alva-Martínez, Fernández, Sarma, & Nandini, 2009); and the development of techniques to remove *Microcystis* and microcystins by coagulation and flocculation of organic matter (Sandoval-Reyes & Ramírez-Zamora, 2019).

Restoration methods

The importance of determining the underlying causes of poor water quality and resolving them through a combination of methodologies to control external and internal nutrient loading is evident in the successful instances of lake restoration impaired by CyanoHABs. (Lürling & Van Oosterhout, 2013; Ibelings, Bormans, Fastner, & Visser, 2016; Lürling, Waajen, Engels, & Van Oosterhout, 2017).

Preventing or controlling eutrophication is the most sustainable approach to reduce cyanobacteria, and at least three strategies have been considered to decrease eutrophication conditions in a system. The chemical methodology includes artificial aeration and flocculation or precipitation of nutrients by adding aluminum and calcium salts or commercial products such as Phoslock (bentonite clay) (Visser *et al.*, 2016b). The physical technique decreases water residence time and removes nutrient-rich sediments (dredging) (Wang *et al.*, 2018; Waajen, Lürling, & Van de Sande, 2019). Manipulation of the trophic cascade by top-down management is one of the most acknowledged paradigms in aquatic ecology (Carpenter *et al.*, 1987; Lampert & Sommer, 2007), this being widely used regulation processes in temperate systems where the effect of removing planktivorous fish, decreasing predation and allowing the recovery of filter-feeding herbivore populations with the ability to consume cyanobacteria has been evaluated (Triest, Stiers, & Van Onsem, 2016).

However, several assessments demonstrated the low effectiveness of top-down control in aquatic ecosystems with dense CyanoHAB (Gliwicz, 1990; Dickman, Newell, González, & Vanni, 2008; Rondel *et al.*, 2008; Lacerot, Kruk, Lürling, & Scheffer, 2013). Even, suggesting that zooplankton facilitate cyanobacterial dominance by eliminating eukaryotic competitors (Lynch & Shapiro, 1981; Mitra & Flynn, 2006; Ger, Naus-Wiezer, De-Meester, & Lürling, 2019). On the other hand, it has been classified as a success in systems with a strong relationship between primary producers and zooplankton (Havens *et al.*, 1996; Jeppesen *et al.*, 2012; Urrutia-Cordero, Ekvall, & Hansson, 2016). Because of their generalist feeding habits, cladocerans like *Daphnia magna* are considered

effective models in controlling cyanobacteria; however, this species is restricted to temperate climate zones. Other crustaceans such as *Simocephalus mixtus* and *Hyalella azteca* show a great capacity to consume cyanobacterial cells at feeding rates higher than those obtained with a eukaryotic primary producer and in combination with the zooplanktivorous fish remotion would be a model for cyanoprokaryote control (Figueroa-Sánchez *et al.*, 2019).

Conclusions

The dominance of CyanoHABs in Mexican reservoirs has been well documented during the last decades through limnological studies or research on phytoplankton diversity analysis (Gaytan-Herrera, Martinez-Almeida, Oliva-Martinez, Duran-Diaz, & Ramirez-Garcia, 2011; Oliva-Martínez, Godínez-Ortega, & Zuñiga-Ramos, 2014). Experimental ecology since the 1990s showed that certain rotifer and cladoceran species can feed and grow on mixed or pure cyanobacterial diets (Alva-Martínez *et al.*, 2007a; Alva-Martínez *et al.*, 2009). Likewise, some novel species such as amphipods and ostracods were studied, quantifying their filtration rates, and assessing their ability to decrease cyanobacterial cells from the medium in laboratory experiments (Fernández, Nandini, Sarma, & Castellanos-Páez, 2016; Figueroa-Sánchez *et al.*, 2019). Several studies

indicate that continuous predation pressure from zooplanktivorous fish in the tropics significantly reduces herbivory of large cladocerans. Current studies also show the accumulation of cyanotoxins in various trophic levels, including food supplies such as snails and fish, which represents a significant risk to human health, especially when are eaten whole without removing the digestive tract where most cyanotoxins accumulate (Berry & Lind, 2010a; Zamora-Barrios *et al.*, 2019). It is well established that many waterbodies in Mexico are supplied with or receive partially treated wastewater. However, limited research has been done on nutrient control in the watershed. CyanoHAB control actions on a large scale have also been rarely examined (Waajen *et al.*, 2019). Likewise, cyanobacteria in saline and sodic lakes have been evaluated in select waterbodies, but further research is needed across the country. We hope that the scientific information contained in this review will be helpful to promote a management model that follows the water safety guidelines proposed by the World Health Organization.

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